



Methodology: Estimating the cost of air pollution in world cities (2020)

Introduction

The **Cost Estimator** is an online tool that estimates the real-time health impact and economic cost from fine particulate matter (PM_{2.5}) air pollution in major world cities.¹ It is deployed in a collaboration between Greenpeace Southeast Asia, IQAir and the Centre for Research on Energy and Clean Air (CREA). Estimates of real time health and economic costs can be viewed on the [Cost Estimator](#) website.²

Methodology & input data

The Cost Estimator is based on methodology used for the Greenpeace Southeast Asia & CREA 2020 [Toxic Air](#) report, which estimated the global health impact and economic cost resulting from exposure to fossil fuel air pollution.³ The Cost Estimator uses real-time ground-level PM_{2.5} measurements collated in IQAir's database.⁴ The algorithm applies scientific risk models in combination with population and public health data to estimate the health and economic costs of air pollution exposure. Mortality and cost estimates represent the total impact attributable to PM_{2.5}.

Accuracy

The Cost Estimator results represent a useful estimate of the real world cost based on currently available scientific knowledge. Like all research, the data and method used by the Cost Estimator contain uncertainty. This uncertainty relates to the precision and representativeness of pollution measurement data, demographic and epidemiological data and information about the economic cost of health impacts. Scientific understanding of the relation between exposure to air pollution and the associated risks for health is an active area of research; the latest available knowledge has been used. In case of missing data, conservative assumptions have been adopted, such that there is an inbuilt tendency towards underestimation of the cost of air pollution. Specifically the Cost Estimator accounts only for the impact of PM_{2.5}, and not for other air pollutants which are also known to have negative health impacts.

¹ Myllyvirta, L. (2020) Revealing the Cost of Air Pollution in World's Cities – in Real Time. July 2020. CREA. <https://energyandcleanair.org/revealing-the-cost-of-air-pollution-in-real-time/>

² <https://www.greenpeace.org/international/campaign/tracking-cost-air-pollution/>

³ Farrow, A., Miller, K. A. & Myllyvirta, L. (2020) Toxic air: The price of fossil fuels. Seoul: Greenpeace Southeast Asia. February 2020. Greenpeace Research Laboratories Technical Report 02-2020: 44 pp

⁴ either from direct measurement or derived from PM₁₀ measurement.

Step by step methodology

1) Air Pollution Data

Sensor types

The Cost Estimator uses measurements of $PM_{2.5}$ from ground-based outdoor pollution monitors in cities around the world. The devices are deployed and operated by governments, organizations such as PurpleAir, and private individuals. The data is sourced and aggregated continuously by IQAir.⁵

City-wide value

Most cities have several monitors, but the Cost Estimator uses one air pollution value for the entire city. Where there are multiple monitors within a city, their median value is used. Pollution is recorded hourly. The interval for computing 24-hour averages starts at 0:00 UTC in all cities around the world.

Quality controls

Monitors with non-governmental operators undergo quality checks by IQAir before being incorporated into the database. The operator demonstrates by photo submission that the sensor is in a suitable location, which means outdoors and sheltered from precipitation, wind and direct sunlight. In addition, IQAir continuously checks that the carbon-dioxide (CO_2) concentration, humidity and temperature readings reflect outdoor conditions during operation.

IQAir performs quality assurance checks on the aggregated data, for example by removing outlying data points when compared with neighbouring monitoring stations. The measurements for each monitoring station are grouped according to the city in which they are located.

Indirect $PM_{2.5}$ data

Some monitors only measure coarse particulate matter (PM_{10}) and do not measure $PM_{2.5}$. For these monitors, the $PM_{2.5}$ values are derived from the PM_{10} measurement using a machine learning system. This system introduced an additional uncertainty; therefore only those cities where the share of native $PM_{2.5}$ monitors is greater than 90% are included.

Representativeness / city filtering

It is assumed by the Cost Estimator that the median of annual mean $PM_{2.5}$ concentrations measured by ground level monitoring sites is representative of exposure in each city. However, cities are heterogeneous and this assumption introduces a significant uncertainty. Monitors in each city were assessed to identify obvious bias that could lead to overestimation of $PM_{2.5}$ exposure. The station locations were overlaid on satellite imagery and manually checked through visual inspection, including the use of street level photography where available. Cities are only included where (a) the monitors show a reasonable geographic spread across the metropolitan area (no clustering, even distribution

⁵ IQAir. "2019 World Air Quality Report." (2020).



between centre, city domain and suburbs) and (b) less than 50% of the monitors are closer than 50 metres to major roads. These criteria provide a practical method for screening any city where monitor locations are very likely to cause an overestimation of $PM_{2.5}$ concentration. As their application is necessarily subjective to some degree; this process was independently carried out by two different analysts.

Data completeness

Real-world air pollution datasets are rarely complete. Outages during maintenance and technical faults result in missing data. On days without pollution measurement for a city, the previous day's measurement is used as a best guess. Only cities with >75% data availability over the year are included.

Other pollutants

Besides $PM_{2.5}$, there are other pollutants which are known to cause damage to human health, such as nitrogen dioxide (NO_2), ozone (O_3) and sulfur-dioxide (SO_2). Although many of the cities included in the Cost Estimator also have monitors for these pollutants, data availability is generally much poorer. Furthermore, health impacts from NO_2 are not included in our results because NO_2 has a high spatial variability and thus the data from point measurements are less representative for the whole city.

2) Demographic and Epidemiological Data

To assess the health impacts, the Cost Estimator uses population and epidemiological data for the metropolitan area of each city. The Cost Estimator assumes that the measured pollutant concentrations aggregated by IQAir provide a reasonable representation of pollution exposure across the metropolitan area of each city (see previous section). Population numbers for most world cities are taken from the United Nations World Population Prospects 2019,⁶ except for Taiwanese cities (Dept. of Household Registration, Ministry of the Interior, Taiwan⁷), Nairobi (Kenya Bureau of Statistics⁸), Delhi and Patna (India Population 2020⁹ and India Population 2019¹⁰) and Turkish cities (Turkish Statistical Institute¹¹).

The data sources of the cities' incidence rates of each health impact as well as the risk functions used to compute the fraction of the health impacts that is attributable to air pollution are listed in Table 1.

⁶ United Nations, Department of Economic and Social Affairs, Population Division (2019). World Population Prospects 2019: Volume I: Comprehensive Tables.

⁷ Dept. of Household Registration, Ministry of the Interior, Taiwan. (2020). <https://gis.ris.gov.tw/dashboard.html>

⁸ Kenya Bureau of Statistics (2019) 2019 Kenya Population and Housing Census Volume I: Population by County and Sub-County

⁹ India Population of 2019. (2019) <https://indiapopulation2019.com/population-of-patna-2019.html>

¹⁰ India Population of 2020. (2020) <https://indiapopulation2020.in/population-of-delhi-2020.html>

¹¹ Turkish Statistical Institute. (2020) <https://www.tuik.gov.tr/>

Table 2. Relative risk at a 10 µg/m³ increase in concentration for various health impacts and data source of the incidence rates.^{12,13,14,15,16,17,18,19,20,21,22}

Health impact	Pollutant	Metric	Relative risk $r(x-x_0=10 \mu\text{g}/\text{m}^3)$	Detection threshold $x_0 (\mu\text{g}/\text{m}^3)$	Relative Risk Source	Incidence Rate Source ²³
Child asthma emergency room visits ²⁴	PM _{2.5}	Annual mean	1.025	6	Zheng et al. 2015	Anenberg et al 2018
Adult asthma emergency room visits ²⁵	PM _{2.5}	Annual mean	1.023	6	Zheng et al. 2015	Anenberg et al 2018
Preterm births	PM _{2.5}	Annual mean	1.15	8.8	Trasande et al 2016	Chawanpaiboon et al 2019
Days of work absence	PM _{2.5}	Annual mean	1.039	0	WHO HRAPIE 2013	EEA 2014
Adult deaths, years of life lost and years of life with disability	PM _{2.5}	Annual mean	non-linear	2.4	Burnett et al ²⁶	GBD 2017

¹² Zheng, Xue-yan, et al. (2015). Association between air pollutants and asthma emergency room visits and hospital admissions in time series studies: a systematic review and meta-analysis." PloS one 10.9. <https://doi.org/10.1371/journal.pone.0138146>

¹³ Trasande L, et al. (2016). Particulate Matter Exposure and Preterm Birth: Estimates of U.S. Attributable Burden and Economic Costs. *Environmental Health Perspectives* 124:12. <https://doi.org/10.1289/ehp.1510810>

¹⁴ Dadvand P et al. (2013). Maternal Exposure to Particulate Air Pollution and Term Birth Weight: A Multi-Country Evaluation of Effect and Heterogeneity. *Environmental Health Perspectives*. https://ehp.niehs.nih.gov/doi/full/10.1289/ehp.1205575?url_ver=Z39.88-2003&rfr_id=ori:rid:crossref.org&rfr_dat=cr_pub%3dpubmed

¹⁵ Malley CS, et al. (2017). Updated Global Estimates of Respiratory Mortality in Adults ≥30Years of Age Attributable to Long-Term Ozone Exposure. *Environmental Health Perspectives* 125:8 CID: 087021 <https://doi.org/10.1289/EHP1390>

¹⁶ World Health Organization. (2013) Health risks of air pollution in Europe—HRAPIE project recommendations for concentration–response functions for cost–benefit analysis of particulate matter, ozone and nitrogen dioxide. UN City: Copenhagen, Denmark

¹⁷ Achakulwisut, Pattanun, et al. (2019). Global, national, and urban burdens of paediatric asthma incidence attributable to ambient NO₂ pollution: estimates from global datasets. *The Lancet Planetary Health* 3.4 (2019): e166-e178.

¹⁸ Anenberg, Susan C., et al. (2018). Estimates of the Global Burden of Ambient PM_{2.5}, Ozone, and NO₂ on Asthma Incidence and Emergency Room Visits. *Environmental health perspectives* 126.10: 107004. <https://doi.org/10.1289/EHP3766>

¹⁹ Chawanpaiboon S, et al. (2019) Global, regional, and national estimates of levels of preterm birth in 2014: a systematic review and modelling analysis. *Lancet Glob Health* 7(1):e37-e46. doi: 10.1016/S2214-109X(18)30451-0

²⁰ World Bank

<https://datahelpdesk.worldbank.org/knowledgebase/articles/906519-world-bank-country-and-lending-groups>

²¹ GBD 2017 Mortality Collaborators. (2018). Global, regional, and national age-sex-specific mortality and life expectancy, 1950–2017: a systematic analysis for the Global Burden of Disease Study 2017. *The Lancet*. 392:1684-735.

²² European Environment Agency (EEA) (2014) Costs of air pollution from European industrial facilities 2008–2012 — an updated assessment. EEA Technical report No 20/2014.

<https://www.eea.europa.eu/publications/costs-of-air-pollution-2008-2012>

²³ These national-level data sources are used when city-level data is not available. When data for a country is not included in the data, the median value for all countries is used.

²⁴ x_0 from: Anenberg et al., *Estimates of the Global Burden of Ambient PM_{2.5}, Ozone, and NO₂ on Asthma Incidence and Emergency Room Visits*, *Environmental Health Perspectives* 126:10 CID: 107004 <https://doi.org/10.1289/EHP3766>

²⁵ idem

²⁶ Burnett R, et al. (2018) Global estimates of mortality associated with long-term exposure to outdoor fine particulate matter. *Proceedings of the National Academy of Sciences*, 115 (38) 9592-9597

3) Health Impact Assessment

The Cost Estimator links specific health impacts to air pollution using scientific risk models from the literature for each impact. The health impacts listed in Table 1 are included in the estimator.

The calculation applies concentration-response functions to the pollutant concentrations of PM_{2.5} to determine the portion of the city's incidence/prevalence rates of different health impacts caused by air pollution. A concentration-response function relates a pollutant concentration to the impact of that pollutant on a population.

Deaths, years of life lost and years lived with disability due to PM_{2.5} exposure are calculated using non-linear, age-specific risk, which gives the increase in risk of different health effects as a function of pollutant concentration, compared with clean air (see Table 1 for references).

For these health impacts, the annual incidence m of a health impact in a specific city is related to the city's population P , the cause-specific death rate in the country m_0 , the risk ratio r and the annual average pollutant concentration x_a by

$$m = P \cdot m_0 \cdot (1 / r(x_a)),$$

where $r(x_a)$ is an empirical function (adopted from the literature). For most health impacts, $r(x_a)$ is modelled by a log-linear risk function, meaning that r increases in a multiplicative manner per incremental increase in pollutant concentration:

$$r(x_a) = \exp[c \cdot (x_a - x_0)] \quad \text{if } x_a > x_0 \qquad r = 1 \text{ otherwise,}$$

with c being an empirical constant (also called "concentration-response factor") and x_0 the *detection threshold* (also called "no-risk threshold", Table 1 and Figure 1), which is usually the lowest concentration found in the study population - it is based on absence of direct evidence, not evidence that there is no risk. Excluding impacts below this threshold is a very conservative assumption.

For presentation of real-time data on the [Cost Estimator](#) website these annual costs are then apportioned linearly to the period of the calendar year that has elapsed in proportion to the average pollutant concentrations of the year to date, such that time periods with higher pollution incur greater costs. The year-to-date incidence m_d is calculated from the past-365-day incidence by

$$m_d = m \cdot [(x_d - x_0) / (x_a - x_0)] \cdot t / 365d \quad \text{if } x_d, x_a > x_0; \quad 0 \text{ otherwise,}$$

where x_d is the year-to-date average pollutant concentration and t is the elapsed time in the current calendar year.

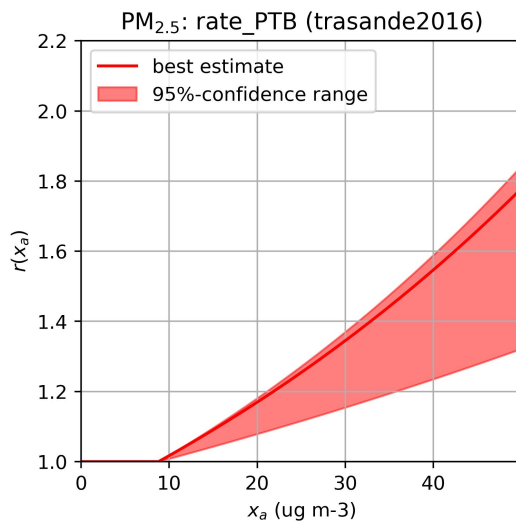


Figure 1: Relative risk function $r(x_a)$ for preterm births due to $PM_{2.5}$ pollution (Trasande et al. 2016). Note the discontinuity of the central/best estimate at the detection threshold $x_0 = 8.8 \mu\text{g}/\text{m}^3$ as a result of the conservative assumption $r(x_a < x_0) = 1$.

Cost Assessment

Health impacts associated with air pollution exposure have economic consequences. Principal among these are direct health care costs, (for example the cost of asthma treatment,) and the cost of reduced economic productivity. Economic productivity may be reduced by absence from or reduced capacity for work.

In addition, a higher risk of death or disability affects people’s well-being; this welfare cost is measured in money by converting it to the amount that people are willing to pay in order to avoid the negative health outcome (“willingness-to-pay”).

These two different types of cost are included where possible and are combined to a total cost. The health impacts listed in Table 2 are included in the calculation as described below:

Lost life years

The economic losses from these air pollution-related deaths are assessed based on the resulting reduction in life expectancy, with one year of life lost equating to economic loss of EUR 56,000 (USD 67,000) in the European Union, following the EEA cost-benefit methodology²⁷, and adjusted by Gross

²⁷ European Environment Agency (EEA) (2014) Costs of air pollution from European industrial facilities 2008–2012 — an updated assessment. EEA Technical report No 20/2014. <https://www.eea.europa.eu/publications/costs-of-air-pollution-2008-2012>



National Income (GNI) Purchasing Power Parity (PPP) by country or region, with an elasticity of 0.9 as recommended by Organisation for Economic Co-operation and Development (OECD)²⁸.

Diabetes, asthma and other chronic respiratory diseases, and disabilities caused by stroke

The Global Burden of Disease project²⁹ has quantified the degree of disability caused by each disease into a “disability weight” that can be used to compare the costs of different illnesses.

The economic cost of disability and reduced quality of life caused by diabetes and chronic bronchitis is assessed based on these disability weights, combined with the economic valuation of disability used by the UK environmental regulator DEFRA³⁰. The valuation is adjusted by GNI PPP for other countries. For example, type 2 diabetes without complications has a disability weight of 4.9%, meaning that the cost of one year lived with diabetes is estimated at 4.9% of the cost of one year lived with disability, or USD 4,000 in the UK and USD 1,600 at world average income level.

The economic cost of asthma related to air pollution is assessed based on two indicators: new cases of asthma linked to NO₂ exposure, and emergency room visits related to PM_{2.5} exposure. An assessment of the direct and indirect cost per year associated with childhood asthma, including medical costs and loss of income to the child’s caregiver, found a cost of USD 3,800 and USD 4,000 in two different communities in California, United States³¹. The midpoint of these two valuations is used for the estimates, adjusted by the ratio of California’s Gross Regional Product to the US national average, and by GNI PPP for other countries. The cost of an emergency room visit is taken from the same study.

Exposure to PM_{2.5} is very likely linked to an even larger number of new asthma cases globally than exposure to NO₂, but uncertainty in the estimates is large³², so this effect is not included. Instead, only the economic cost of emergency room visits for asthma linked to PM_{2.5} exposure are included. This is only a small part of the overall cost of the burden of asthma linked to PM_{2.5}.

Preterm Birth and Low Birth Weight

PM_{2.5} exposure to pregnant women increases the likelihood of preterm birth and low birth weight, both of which in turn increase the risk of many health and development issues throughout a baby’s life. An estimated economic cost of a preterm birth, which primarily results from lower economic productivity

²⁸ OECD. (2016). The Economic Consequences of Outdoor Air Pollution, OECD Publishing, Paris. <http://dx.doi.org/10.1787/9789264257474-en>

²⁹ Cohen, A. J. et al. (2017). Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: an analysis of data from the Global Burden of Diseases Study 2015. *Lancet* 389, 1907–1918

³⁰ Birchby D, et al. (2019). Air Quality damage cost update 2019. Report for Defra. AQ0650. Ricardo Energy & Environment, United Kingdom.

³¹ Brandt SJ, et al. (2012) Costs of childhood asthma due to traffic-related pollution in two California communities. *European Respiratory Journal* Aug 2012, 40 (2) 363-370

³² Anenberg, Susan C., et al. (2018) Estimates of the Global Burden of Ambient PM_{2.5}, Ozone, and NO₂ on Asthma Incidence and Emergency Room Visits. *Environmental health perspectives* 126.10: 107004. (2018).



and increased health care costs, of USD 300,000 per birth in the US is used.³³ The valuations are adjusted using GDP PPP within regional groups.

Work Absence

Exposure to PM_{2.5} air pollution leads to increased work absences due to illness (sick leave). The incidence of work absences is estimated using WHO recommended concentration response functions³⁴. The economic cost of sick leave is evaluated at USD 145 per day in the European Union, based on EEA recommendations³⁵, and adjusted for other countries based on GDP PPP.

The method and data used are described in further in reports by CREA³⁶ and Greenpeace Southeast Asia³⁷.

³³ Trasande L, et al (2016) Particulate Matter Exposure and Preterm Birth: Estimates of U.S. Attributable Burden and Economic Costs. *Environmental Health Perspectives* 124:12.

³⁴ World Health Organization. (2013) Health risks of air pollution in Europe—HRAPIE project recommendations for concentration–response functions for cost–benefit analysis of particulate matter, ozone and nitrogen dioxide. UN City: Copenhagen, Denmark

³⁵ Converted from EUR to USD. Source: European Environment Agency (EEA) (2014) Costs of air pollution from European industrial facilities 2008–2012 — an updated assessment. EEA Technical report No 20/2014. <https://www.eea.europa.eu/publications/costs-of-air-pollution-2008-2012>

³⁶ Myllyvirta, L. (2020) Revealing the Cost of Air Pollution in World's Cities – in Real Time. July 2020. CREA. <https://energyandcleanair.org/revealing-the-cost-of-air-pollution-in-real-time/>

³⁷ Farrow, A. et al (2020) Toxic air: The price of fossil fuels. Seoul: Greenpeace Southeast Asia. February 2020. Greenpeace Research Laboratories Technical Report 02-2020: 44 pp



Contributors and internal review

IQAir collected, city-aggregated, quality-checked, filtered and managed the measurement data of the on-site air pollution monitors.

CREA conceptualized the health impact assessment & economic cost computation and implemented it in computer code.

Greenpeace conducted a technical review of the methods and their implementation and analysed the resulting health impact and cost data set.